

1991

Sustainable harvesting of tropical rainforests: Reply to Keto, Scott and Olsen

Jerome K. Vanclay
Southern Cross University

E.J. Rudder
Queensland Forest Service,

Glenn Dale
Tree Crop Technologies Pty Ltd

G A. Blake
Queensland Forest Service,

Publication details

Post-print of Vanclay, JK, Rudder, EJ, Dale, G & Blake, GA 1991, 'Sustainable harvesting of tropical rainforests: Reply to Keto, Scott and Olsen', *Journal of Environmental Management*, vol. 33, no. 4, pp. 379-394.

Journal of Environmental Management home page available at www.elsevier.com/locate/jenvman

Publisher's version of article available at [http://dx.doi.org/10.1016/S0301-4797\(05\)80025-8](http://dx.doi.org/10.1016/S0301-4797(05)80025-8)

Sustainable Harvesting of Tropical Rainforests: Reply to Keto, Scott and Olsen

J. K. Vanclay, E. J. Rudder, G. Dale and G. A. Blake

Queensland Forest Service, GPO Box 944, Brisbane, Queensland, Australia. Received 20 August 1990

This paper refutes the Keto *et al.* proposition that the Queensland selection logging system is neither ecologically nor economically sustainable. The key requirements of this system are: (1) that logging guidelines are sympathetic to the silvicultural characteristics of the forest, ensuring adequate regeneration of commercial species and discouraging invasion by weeds; (2) tree-marking by trained staff specifies trees to be retained, trees to be removed and the direction of felling to ensure minimal damage to the residual stand; (3) logging equipment is appropriate and driven by trained operators to ensure minimal damage and soil disturbance, compaction and erosion; (4) prescriptions ensure that adequate stream buffers and steep slopes are excluded from logging; (5) sufficient areas for scientific reference, feature protection and recreation are identified and excluded from logging; and (6) that deficiencies in an evolving system are recognized and remedied, leading to an improved system. Many studies of the effects of logging in these forests have been published and collectively provide a unique demonstration of one possible approach to sustainable timber harvesting.

Keywords: tropical rainforest, sustainability, Queensland, Australia.

1. Introduction

We are all concerned at the rate and extent of tropical deforestation and degradation of forested lands. However, the presentation by Keto *et al.* (1990) may do little to alleviate the problem. There are indications that, with appropriate management, sustainable timber harvesting can be achieved with minimal environmental impact (e.g. Jonkers and Schmidt, 1984; Dawkins, 1988). To try to convince tropical timber producers otherwise is to invite the broadscale conversion of rainforests to other land uses! A more effective way to ensure the conservation of the world's tropical rainforests may be to promote sustainable harvesting of timber and other forest products (Colinvaux, 1989; Stocker, 1989; Vanclay, 1991c). This view is not restricted to foresters and forest services, but is also promoted by some conservation groups (Thompson, 1988).

Keto *et al.*'s (1990) critique of the Queensland model of sustainable timber production is largely restricted to questioning the validity of the permanent sample plots which form the basis for timber yield estimates. They suggest that the basis of the model is harvesting an amount equivalent to the growth between harvests. Whilst such harvesting may be necessary to ensure sustainability, it is not sufficient. Prediction and harvesting of a sustained yield is only part of a sustainable timber production system. The main requirement of the Queensland sustainable timber production system is that the forest is left in "good" condition, and this requires that:

1. Logging guidelines are sympathetic to the silvicultural characteristics of the forest, viz. ensuring retention of vigorous advance growth, harvesting only defective and mature trees, providing for adequate regeneration of commercial species and discouraging invasion by weeds, such as bamboo and climbing vines.
2. Treemarking by trained staff specifies trees to be retained, trees to be removed and the direction of felling to ensure minimal damage to growing stock and minimal opening of the canopy.
3. Logging equipment is appropriate and driven by trained operators to ensure minimal damage to the residual stand and minimal soil disturbance, compaction and erosion.
4. Prescriptions ensure that adequate stream buffers and steep slopes are excluded from logging.
5. Sufficient areas for scientific reference, feature protection and recreation are identified and excluded from logging.
6. Deficiencies in an evolving system are recognized and remedied, leading to an improved system.

Provided that these principles are adhered to and the forest is left in good condition, it may not matter if

the sustained yield is exceeded for a short time. In Queensland, it was government policy to stimulate industrial development by exceeding the sustained yield during harvesting of the virgin resource, but it was realized as early as 1949 that the harvest would ultimately need to be reduced. Throughout the period 1948-1978, the allocation was set at 207 000 cubic metres per annum, and was progressively reduced to the estimated sustainable yield of 60 000 m³/annum in 1986. Yet, despite this apparent overcutting, the region was still considered worthy of inclusion on the World Heritage List in 1988. Clearly, this harvest (207 000 m³/annum) could only be maintained for so long because it was the first harvest from virgin stands. This gradual introduction of sustainable harvesting in north Queensland parallels the experience in North America (Clawson and Sedjo, 1983; Parry *et al.*, 1983).

Contrary to Keto *et al.*'s (1990) claims, the second harvest had commenced in north Queensland: 28 941 ha of forest previously logged under "cutter selection" to specified girth limits during 1939-1955 were relogged to the treemarking guidelines prior to the 1988 logging ban and yielded economically viable harvests. Whilst some of this harvest came from species previously considered less desirable, or from areas simply missed or passed over during the first harvest (logging under cutter selection was typically very selective and restricted to easily accessible areas), some of the harvested volume may be attributed to actual growth on trees which were too small at the time of first harvest.

Simple arithmetic demonstrates the feasibility of the sustained yield. A yield of 60 000 m³/annum from 160 000 ha (Preston and Vanclay, 1988) implies an average annual increment of only 0.375 m³/ha/annum, which is a reasonably conservative estimate. Assuming a 40-year nominal rotation implies that 4000 ha would be logged annually, and that the average yield per hectare would be 15 m³/ha. This is a realizable volume consistent with volumes attained in recent recut areas. However, some recut areas have realized much higher yields (e.g. Beatrice Logging Area averaged 40 m³/ha).

2. Permanent sample plots

The main thrust of Keto *et al.*'s (1990) paper was a criticism of the permanent sample plots maintained by the Queensland Forest Service (QFS) in north Queensland. Whilst it is easy to be critical, it is appropriate to bear in mind that some of these plots were established as early as 1948, before the advent of computers and simulation systems, during an era when there was little doubt that the rainforests should be exploited. It is not easy to foresee, at plot establishment, all the possible uses to which plot data may be put, and their exact measurement requirements. Indeed, Whitmore (1989) records that he had to abandon one of his projects in the Solomon Islands because the plots he established in 1964 did not record sufficient detail. In this light, the QFS database of 247 plots (see Appendix 1) has shown stability and versatility. Most plots have been measured every 5 years (sometimes more frequently) for up to 40 years, with only one change in measurement procedure (in 1981 the minimum size for inclusion was changed from 6 m height to 10 cm diameter). All trees have been individually numbered and tagged so that the development of each individual tree could be reliably traced. Whilst this is essentially simple, it involves a huge amount of data, a considerable budget and dedication and diligence by field and office staff.

Keto *et al.* (1990) conveyed the impression that yield estimates for Queensland's rainforests were derived by estimating average plot volume growth and extrapolating this increment to the whole forest estate. Were this the case, it would be necessary to ensure that plots were typical and representative. In fact, a more sophisticated and flexible methodology has been used, and was described in three of the references quoted by Keto *et al.* (*i.e.* Vanclay, 1983; Preston and Vanclay, 1988; Poore, 1989). This approach employed these permanent sample plots only to develop a growth model, a computer simulation system which predicts growth and change in the rainforest under a wide variety of conditions (e.g. Vanclay 1989a). The present state of the forest is determined from a large number (319 in Preston and Vanclay's study, 518 in more recent unpublished studies) of temporary inventory plots, and yields are determined by repeatedly simulating the growth and harvesting of each of these inventory plots (Vanclay and Preston, 1989). With this growth modelling approach, it is important to sample the widest possible range of stand conditions (Box, 1966; Vanclay, 1991a), not merely the "typical" stands.

Contrary to the claims of Keto *et al.* (1990), records of treatment history and intensity are available for all permanent sample plots (except for the 1929 treatments). Treatment was not "often repeated and of unknown intensity", but applied once, twice and in few instances on three occasions (see Appendix 1). These treated plots were not used in the development of Vanclay's (1989a) growth model or in Preston and Vanclay's (1988) yield calculation. These data have subsequently been used in a revised growth model (Vanclay, 1991b), but do not detract from the utility of that model as it contains an expression to account

for the effects of treatment. Whilst it may be surprising to many readers that so many plots were subjected to treatment and/or underplanting, it should be recognized that, at plot establishment, there was no doubt that the rainforests were to be exploited, and the great research question was how commercial timber production could be increased in a cost-effective way (Henry, 1960).

Rarely has regeneration been unsuccessful in these rainforests. On the contrary, regeneration has been so abundant that it provoked a whole series of thinning trials to identify optimal spacings (see Appendix 1). However, underplanting and enrichment planting have been less successful. Red cedar (*Toona australis*) underplantings have been successful only on State Forest 191 from plantings in 1914, and in Experiment 166 from plantings during the 1950s. Hoop pine (*Araucaria cunninghamii*) showed promise as an underplant only with regular weed control to eliminate competition. *Flindersia* species have also shown little promise (Keys, 1979). Planted trees have been so identified in the data, and statistical analyses revealed no significant difference in growth rate (compared to natural regeneration) once trees had attained 10 cm diameter. Although enrichment planting on permanent sample plots may alter species composition of these plots, it can have no influence on yield estimates, as individual species are identified on the permanent sample plots, in the growth simulation model, and in the inventory plots used in calculations (Vanclay and Preston, 1989).

Whilst prism plots sample only arboreal vegetation and thus provide limited utility for detailed ecological studies, there is no reason to doubt their efficacy in providing growth data of forest trees (Myers and Beers, 1968). The method is highly efficient in estimating variables such as stand basal area and volumes, and is the most efficient way to enumerate tree frequencies by diameter class in tropical forests (Schreuder *et al.*, 1987). The eight prism plots in question were initially established to investigate the effects of logging on the residual stand, and have fulfilled that purpose adequately (Vanclay, 1989b). Keto *et al.* (1990) argue that most of the remaining permanent plots should be discarded because they are less than 0.4 ha. Certainly, it is preferable that plots should have a standard size (ideally 0.4-0.5 ha), but smaller plots are in no way invalid, and still contribute useful growth information. Larger plots may be impractical, as "it is difficult to find many sites in the region larger than 0.5 ha which do not include major physical or floristic discontinuities" (West *et al.* 1988).

Keto *et al.* (1990) contend that several of the unlogged plots are not representative of virgin rainforest, but represent successional communities dominated by secondary species (e.g. *Acacia aulacocarpa*). However, *A. aulacocarpa* has never been recorded on Plot 626/2 (prior to 1981 all stems exceeding 6 m height were measured, but since 1981 only stems exceeding 10 cm diameter have been measured), although it is abundant on roadsides in the vicinity of the plot. Similarly, no *A. aulacocarpa* has ever been recorded on the virgin plot at Mt Windsor (679/2). This plot contains several trees over a metre in diameter, and the largest trees exceed 135 cm. Large trees in this plot include *Cardwellia sublimis*, *Ceratopetalum succirubrum*, *Flindersia pimenteliana*, *Planchonella papyracea* and *Syzygium wesa*, which are fairly typical of these granite soils.

Keto *et al.* (1990) rejected several plots because they had no large commercial stems. However, one characteristic of Culpa Lands (south of Koombooloomba) where some of these plots are located, is the absence of large trees. Thus these plots may well be typical of a considerable area. Not all rainforest has big trees!

Many of the plots were originally half-acre (0.2023 ha) plots measuring two chains by two and a half chains (c. 40 x 50 m) or one chain by five chains (c. 20 x 100 m). Thus, the great majority of plots are 40 m or less in width. Keto *et al.* (1990) reject as "too small (0.12-0.15 ha) and/or too narrow (10-20 m)" several plots (e.g. 608/1, 610/1, 623/1) which are exactly the same size and dimensions as the two plots (612/1, 613/1) accepted by Keto *et al.* as "usable".

Keto *et al.* (1990) also rejected many plots which "were reduced in effective area by roads, snig tracks, creeks and impeded drainage" or by granite boulders and landslips. Large granite boulders, landslips and cyclones are all natural phenomena in north Queensland, as in most tropical moist forests, and should be represented in an unbiased sample. Similarly, snig tracks, creeks and areas of impeded drainage are common phenomena and should be included in samples. Any system of permanent plots which failed to sample these phenomena could be accused of subjective bias. No permanent sample plots include permanent roads, although some may have been traversed by temporary logging extraction tracks. In any case, data collected prior to disturbance (landslip, cyclone, inundation) is not lost, and continues to provide suitable predisturbance baseline data.

Keto *et al.* (1990) reject several plots which have fewer than 16 years of measurement, claiming that these are of "dubious statistical validity". Whilst such short measurement histories will not enable the

detection of subtle long-term trends, they still provide good growth data, and there is no reason to doubt their statistical validity simply because of their relatively short history. In contrast, some biometricians argue that there are statistical gains to be attained by measuring plots for a few years only, before abandoning these and establishing new plots elsewhere (e.g. Tennent, 1988).

Keto *et al.* (1990) conclude their criticism of plots with a quote from Vanclay (1983, p. 161) which suggested that available data were often inadequate for detailed growth modelling studies. However, that quote is out of context. Vanclay was not referring to the rainforests of north Queensland, nor to the then Queensland Department of Forestry, but commenting on the difficulties generally facing modellers of indigenous forests everywhere. At that time (1983), Vanclay had no first-hand knowledge of the data from north Queensland.

We make no claim that the QFS has a perfect database for all timber yield and ecological studies, but few resource managers are lucky enough to have complete and perfect information. The art of land use planning and resource management is to make the best possible use of incomplete and imperfect information. The QFS database is deficient in increment data for *Backhousia bancroftii*, a major commercial species, and the database used to construct the revised growth model (Vanclay and Preston, 1989; Vanclay, 1991b) employed data from the CSIRO EP series of plots (West *et al.*, 1988) to overcome this deficiency. We hope that additional plots will be established to extend the present database further. There should be no stigma in admitting a weakness in a database or management system; the very process of improvement requires that deficiencies are recognized and remedied.

3. Ecological sustainability

The effect of disturbance on rainforest structure, species diversity and species richness is controversial, and research findings are very much subject to sample size and degree of disturbance. Whilst Keto *et al.* (1990) suggest that several overseas studies support their contention that repeated logging will lead to a reduction in species diversity and richness, they omit any reference to alternative views and discussions in the literature [e.g. Whitmore's (1984) response to Denslow (1980), and Nicholson *et al.*'s (1990) reply to Saxon (1990)]. Boyce (1988) and Brunig (1988) argue that selection logging actually increases diversity. Crome *et al.* (1990) found no loss of species as a result of logging. In north Queensland, the rare marsupial *Antechinus godmani* has a restricted distribution, but is abundant in an area logged twice and traversed by two major roads. Wyatt-Smith (1988) concluded that "The polycyclic selection logging system of management as currently practiced in northern Queensland rain forest cannot in any way be considered to pose a threat to the continued existence of 'threatened' species of fauna or flora".

An important component of the Queensland selection logging system is the exclusion of logging from scientific areas, feature protection areas, steep slopes and stream buffers, and the effect of this is to create a mosaic of logged and unlogged forest. In addition, logging does not completely destroy the canopy. Guidelines prescribe that not more than 50% of the canopy should be disturbed, and recent studies indicated that 40 to 60% of the area actually designated for logging could remain completely undisturbed (Applegate, 1989). Crome *et al.* (1990) found that less than 25% of the canopy was lost as a result of logging. Rainforests may be more resilient than is popularly believed. One small study (King and Chapman 1983) found that 25 years after clearfelling of all merchantable stems in a *Ceratopetalum-dominated* warm temperate rainforest, all flowering plants, ferns and mosses that were originally present could again be found. In north Queensland, Stocker (1981) found that 82 tree species regenerated within 2 years of felling and burning rainforest. A comprehensive literature review (Horne and Hickey 1991) found that few quantitative studies of the effects of selection logging had been made, but concluded that the environmental impacts may be minor. Baur (1988) concluded that "Whilst more checking and research are necessary, there seem good grounds for believing that the selective logging system, with its mosaic of disturbed and undisturbed patches and with its similarity to the natural processes experienced in the rainforest, represents no threat to the survival of any plant or animal species".

In appraising the impact of logging, it is necessary to specify whether the sample comprises only areas where the canopy was actually removed in logging, or whether it encompasses the adjacent less disturbed area. The former is likely to indicate massive structural changes and a great reduction in diversity and richness. The latter requires a larger sample and is likely to reveal small structural changes and increased diversity and richness. Nicholson *et al.* (1988) also commented on the importance of sample size, and observed that a large sample (2 ha) of logged forest would reveal no loss of species as a result of logging.

The impact of selection logging on these forests has been extensively studied. The effects on fauna

(Crome and Moore, 1989), flora (Nicholson *et al.*, 1988, 1990; Saxon, 1990; Crome *et al.*, in press), hydrology (Gilmour, 1971), soils (Gillman *et al.*, 1985) and timber production (Vanclay and Preston, 1989; Vanclay, 1990) have been studied, and provide no indication that such harvesting is not sustainable.

4. Economic sustainability

The net economic benefit or cost of rainforest logging cannot be estimated by a simple financial examination of QFS revenues and expenditure. Keto *et al.* (1990, table 2) overstate actual QFS expenditure associated with rainforest harvesting by including expenditure on plantation establishment and maintenance. The costs and revenues recorded during the last full year of rainforest logging operations were the only data recorded on a programme basis, and showed a small profit for the rainforest subprogramme.

The rainforest-based forestry and timber industry of north Queensland was an important and economically viable part of the north Queensland economy (ACIL Australia, 1987). Independent studies (Harris, 1987; Cameron McNamara, 1988) identified some 2000 jobs directly or indirectly linked to rainforest logging in north Queensland, whilst value adding by the industry was estimated at \$25 m. per annum (Harris, 1987).

Long-term total economic losses to individuals, industry and Government from the cessation of rainforest logging have been estimated at \$400 m. (Cameron McNamara, 1988, table 1.1). Cameron McNamara (1988) further indicated that lost rainforest exports and imports of replacement products would cost an estimated \$30 m. annually. These costs far outweigh any subsidy provided by the Queensland Government to maintain QFS operations in north Queensland.

5. Conclusion

Keto *et al.* (1990) contend that "future timber supplies can ultimately only come from plantations", but we ask if minimal impact selection logging is not sustainable, how can these more intensive plantations be sustainable? We agree with Keto *et al.* that "protection of remaining forests will be essential", but suggest that production may provide protection for many of these forests (Vanclay, 1991c).

Keto *et al.* (1990) have not fulfilled their stated objective to "examine that model, its deficiencies and the potential for application to developing countries". Rather, they have criticized several specific aspects. The importance of the Queensland example lies in the successful implementation and co-ordination of many components including reliable resource inventory, estimating the sustained yield, determining areas to be logged, planning the required extraction infrastructure, supervising felling and extraction, ensuring adequate erosion controls on completion of logging, and maintaining reliable management records. These practices and principles, which have been developed to satisfy operational requirements in north Queensland, could serve as examples to other tropical countries and have formed the basis for the ITTO guidelines (ITTO, 1990). The recent World Heritage listing of 97% of these tropical rainforests which have been used for timber production for more than a century (and more intensively managed during the past 40 years) is testimony to the standard of management and the success of the Queensland selection logging system.

The data derived from the permanent sample plots in north Queensland can provide no useful information for other tropical countries; they will need their own plots to predict yields and monitor changes. The Queensland permanent sample plots can merely demonstrate a proven methodology for data collection, management and analysis which may be used elsewhere. Queensland foresters are privileged to have over 40 years' experience in the establishment and maintenance, not only of a permanent sample plot system, but of an integrated forest management system.

Keto *et al.* (1990) have not demonstrated the failure of the "north Queensland logging model" to produce a sustainable harvest of timber, and their criticism of permanent sample plots is flawed. Whilst the Queensland selection logging system is not the only means to ensure a sustainable harvest, it remains one of the best demonstrations visible today. Many other examples (Dawkins, 1988) have been lost through changes in land use.

References

- ACIL Australia Pty Ltd (1987). *The Forestry and Timber Industry of North Queensland-Impact of Proposed World Heritage Listing*. Barton; ACT: A.C.I.L. Australia Pty Ltd.
- Applegate, G. B. (1989). *Comparison of Logging Effects on Two Recut Sites in North Queensland Rainforests, Post 1982*. Unpublished report. Brisbane: Queensland Department of Forestry.
- Baur, G. N. (1988). *World Heritage Listing: North Queensland Wet Tropics*. Unpublished report. Brisbane: Queensland Department of Forestry.
- Box, G. E. P. (1966). Use and abuse of regression. *Technometrics* 8, 625-629.
- Boyce, S. G. (1988). *Systematic Culture of Rainforest in Queensland*. Unpublished report. Brisbane: Queensland Department of Forestry.
- Brunig, E. F. (1988). *The Effect of Poly-cyclic Selection Logging in the Wet Tropical Rainforests of North-east Queensland*. Unpublished report. Brisbane: Queensland Department of Forestry.
- Cameron McNamara (1988). *Socio-economic Study: Wet Tropical Rainforests North Queensland*. Brisbane: Cameron McNamara.
- Clawson, M. and Sedjo, R. (1983). History of sustained yield concept and its application to developing countries. In *History of Sustained Yield Forestry: A Symposium* (Steen, H. K., ed.), pp. 3-15. Portland, Oregon: Forest History Society.
- Colinvaux, P. A. (1989). The past and future Amazon. *Scientific American* 260 (5), 102-108.
- Crome, F. H. J. and Moore, L. A. (1989). Display site constancy of bowerbirds and the effects of logging on Mt. Windsor Tableland, North Queensland. *Emu* 89, 47-52.
- Crome, F. H. J., Moore, L. A. and Richards, G. C. (1992). A study of logging damage in upland rainforest in north Queensland. *Forest Ecology and Management* 49: 1-29.
- Dawkins, H. C. (1988). The first century of tropical silviculture: Successes forgotten and failures misunderstood. In *The Future of the Tropical Rain Forest* (McDermott, J. M., ed.), pp. 4-8. Proceedings of Conference at Oxford, 27-28 June 1988. Oxford Forestry Institute.
- Denslow, J. S. (1980). Patterns of species diversity during succession under different disturbance regimes. *Oecologia* 46, 18-21.
- Gillman, G. P., Sinclair, D. F., Knowlton, R. and Keys, M. (1985). The effect on some soil chemical properties of the selective logging of a north Queensland rainforest. *Forest Ecology and Management* 12, 195-214.
- Gilmour, D. A. (1971). The effects of logging on streamflow and sedimentation in a north Queensland rainforest catchment. *Commonwealth Forestry Review* 50, 38-48.
- Harris, P. (1987). *An Analysis of the Socio-economic Impact of the World Heritage Listing of the Rainforests of North Queensland*. Melbourne: Forest Industries Campaign Association.
- Henry, N. B. (1960). The effect of silvicultural treatment on the production from native forests. *Australian Forestry* 24, 30-45.
- Horne, R. and Hickey, J. (1991). Review: Ecological sensitivity of Australian rainforests to selective logging. *Australian Journal of Ecology* 16, 119-129.
- ITTO (1990). *ITTO Guidelines for the Sustainable Management of Natural Tropical Forests*. ITTO Technical Series 5. Yokohama: International Tropical Timber Organization.
- Jonkers, W. B. J. and Schmidt, P. (1984). Ecology and timber production in tropical rainforest in Surinam. *Interciencia* 9, 290-297.
- Keto, A. I., Scott, K. and Olsen, M. F. (1990). Sustainable harvesting of tropical rainforests: A reassessment. Paper presented at Seminar on Sustainable Development of Tropical Forests held in conjunction with 8th Session of International Tropical Timber Council, 16-23 May 1990, Bali, Indonesia. The same paper was given under the title "Environmental impact of forestry operations and implications for sustainable development" at the UNDP Workshop on Environmental Management and Sustainable Development in the South Pacific, Suva, Fiji, 16-20 April 1990.
- Keys, M. G. (1979). *Growth of Hickory Ash as a Plantation Species and as an Underplant in Treated Rainforest at Kuranda, North Queensland*. Research Note No. 25. Queensland Department of Forestry.
- King, G. C. and Chapman, W. S. (1983). Floristic composition and structure of a rainforest area 25 yr after logging. *Australian Journal of Ecology* 8, 415-423.
- Myers, C. C. and Beers, T. W. (1968). Point sampling for forest growth estimation. *Journal of Forestry* 66, 927-929.

- Nicholson, D. I., Henry, N. B. and Rudder, J. (1988). Stand changes in north Queensland rainforests. *Proceedings of the Ecology Society of Australia* **15**, 61-80.
- Nicholson, D. I., Henry, N. B. and Rudder, J. (1990). Reply: Disturbance regimes in north Queensland rainforests: A re-evaluation of their relationship to species richness and diversity. *Australian Journal of Ecology* **15**, 245-246.
- Parry, B. T., Vaux, H. J. and Dennis, N. (1983). Changing conceptions of sustained-yield policy on the national forests. *Journal of Forestry* **81**, 150-154.
- Poore, D. (1989). Queensland, Australia: An approach to successful sustainable management. In *No Timber Without Trees. A Study for ITTO* (Poore, D., Burgess, P., Palmer, J., Rietbergen, S. and Synnott, T., eds), pp. 28-39. London: Earthscan Publications Ltd.
- Preston, R. A. and Vanclay, J. K. (1988). *Calculation of Timber Yields from North Queensland Rainforests*. Technical Paper No. 47. Brisbane: Queensland Department of Forestry.
- Saxon, E. C. (1990). Comment: Disturbance regimes in north Queensland rainforests: A re-evaluation of their relationship to species richness and diversity. *Australian Journal of Ecology* **15**, 241-244.
- Schreuder, H. T., Banyard, S. G. and Brink, G. E. (1987). Comparison of three sampling methods in estimating stand parameters for a tropical forest. *Forest Ecology Management* **21**, 119-127.
- Stocker, G. C. (1981). Regeneration of a north Queensland rain forest following felling and burning. *Biotropica* **13**, 86-92.
- Stocker, G. C. (1989). Alternative tropical rainforest timber harvesting strategies. Discussion paper at UNESCO Research Planning Workshop on Ecological and Economic Sustainability of Tropical Rainforest Management, Paris, 4-6 September 1989.
- Tennent, R. B. (1988). A New Zealand data collection procedure for growth modelling. In *Modelling Trees, Stands and Forests* (Leech, J. W., McMurtrie, R. E., West, P. W., Spencer, R. D. and Spencer, B. M., eds), pp. 112-120. Proceedings of a Workshop in August 1985 at the University of Melbourne. School of Forestry, University of Melbourne, Bulletin No. 5.
- Thomson, K. (1988). The FoE tropical rainforest campaign. In *The Future of the Tropical Rain Forest* (McDermott, J. M., ed.), pp. 58-61. Proceedings of Conference at Oxford, 27-28 June 1988. Oxford Forestry Institute.
- Tracey, J. G. and Webb, L. J. (1976). *Key to the Vegetation of the Humid Tropical Region of North Queensland*. Map Series. Atherton, Queensland: Rain Forest Ecology Unit, CSIRO.
- Vanclay, J. K. (1983). Techniques for modelling timber yield from indigenous forests with special reference to Queensland. M.Sc. Thesis, University of Oxford.
- Vanclay, J. K. (1989a). A growth model for north Queensland rainforests. *Forest Ecology and Management* **27**, 245-271.
- Vanclay, J. K. (1989b). Modelling selection harvesting in tropical rain forests. *Journal of Tropical Forest Science* **1**, 280-294.
- Vanclay, J. K. (1989c). Site productivity assessment in rainforests: an objective approach using indicator species. In *Proceedings of the Seminar on Growth and Yield in Tropical Mixed/Moist Forests, 20-24 June 1988, Kuala Lumpur* (Wan Razali Mohd., Chan, H. T. and Appanah, S., eds), pp. 225-241. Forest Research Institute, Malaysia.
- Vanclay, J. K. (1990). Effects of selection logging on rainforest productivity. *Australian Forestry*. **53**, 200-214.
- Vanclay, J. K. (1991a). Review: Data requirements for developing growth models for tropical moist forests. *Commonwealth Forestry Review* **70**, 105-111.
- Vanclay, J. K. (1991b). Aggregating tree species to develop diameter increment equations for tropical rainforests. *Forest Ecology and Management* **42**, 143-168.
- Vanclay, J. K. (1991c). Research needs for sustainable forest resources. In *Tropical Rainforest Research in Australia*. (N. Goudberg and M. Bonell, eds), pp. 133-143: Townsville: Institute of Tropical Rainforest Studies, James Cook University.
- Vanclay, J. K. and Preston, R. A. (1989). Sustainable timber harvesting in the rainforests of northern Queensland. In *Forest Planning for People*, pp. 181-191. Proceedings of 13th biennial conference of the Institute of Foresters of Australia, Leura, N.S.W., 18-22 September 1989. Sydney: Institute of Foresters of Australia.
- West, P. W., Stocker, G. C. and Unwin, G. L. (1988). Environmental relationships and floristic and structural change in some unlogged tropical rainforest plots of north Queensland. *Proceedings of the Ecological Society of Australia* **15**, 49-60.

- Whitmore, T. C. (1984). Gap size and species richness in tropical rain forests. *Biotropica* **16**, 239.
- Whitmore, T. C. (1989). Guidelines to avoid remeasurement problems in permanent sample plots in tropical rain forests. *Biotropica* **21**, 282-283.
- Wyatt-Smith, J. (1988). Queensland rain forest: World heritage listing. Unpublished report. Brisbane: Queensland Department of Forestry.

Appendix 1

The following is a list of permanent sample plots in the Queensland Forest Service rainforest database. Geological types are Alluvial (AL), Acid Volcanic (AC), Basic Volcanic (BV), Coarse-grained Granite (CG), Sedimentary and Metamorphic (SM) and Tully fine-grained Granite (TG). Site quality was determined using Vanclay's (1989c) equation 13. Rainforest structural types follow Tracey and Webb (1976). Brief descriptions of the origin of the various plot types are given below.

Notes.

1. Paired treatment plots comparing growth with and without silvicultural treatment.
2. Plots monitoring the development of regeneration.
3. Experiments monitoring development of enrichment plantings following thinning to various spacings.
4. Experiments monitoring development of rainforest following application of different silvicultural treatment prescriptions.
5. Experiments monitoring development of enrichment plantings.
6. Experiment examining benefits of silvicultural treatment 10 years prior to logging, with a view to getting more regeneration.
7. Experiments monitoring effects of retreatment 15 years after initial silvicultural treatment.
8. Treatment of unproductive rainforest attempting to produce a viable timber harvest.
9. Logging damage studies.
10. Plots monitoring development of dense stands of rainforest.
11. Plots monitoring growth and yield in rainforest under routine management. These plots were deliberately located to sample good, average and poor rainforest.
12. CSIRO growth monitoring plots described in West *et al.* (1988).

Expt No.	Plot No.	State Forest	UTM Grid Ref.			Area (ha)	First measure	Last measure	Geol. type	Site quality	Alt. (m)	Aspect	Slope (deg.)	Rain (mm)	Struct. type	Years logged	Years treated	Plot type
			Zone	East	North													
625	2	185	55	350510	8107810	0-2023	69	84	CG	8-8	945	S	25	1650	8/9		11	
625	3	185	55	352310	8108900	0-2023	68	84	CG	8-8	1065	SW	25	1650	9		11	
625	4	185	55	351200	8107360	0-2023	68	84	CG	6-9	790	E	15	1650	8/9	65	11	
625	5	185	55	351150	8106620	0-2023	68	84	CG	7-2	730	SE	20	1650	8	70	11	
626	1	1229	55	355500	8143090	0-2023	69	87	SM	5-4	360	SSW	5	2030	2a	54	11	
626	2	1229	55	354760	8143540	0-2023	69	87	SM	7-0	360	NE	10	2030	2a		11	
640	1	756	55	361100	8052500	PRISM	79	86	BV	—	720	N	15	2000		62, 80	9	
679	1	144	55	290900	8201600	PRISM	80	85	CG	—	1100	N	5	1036		80	9	
679	2	144	55	290600	8201700	PRISM	80	85	CG	—	1060	NE	5	1036			9	
701	1	756	55	371650	8047300	PRISM	85	88	CG	—	400	—		2000		87	9	
EP	2	185	55	349550	8103510	0-5000	71	87	CG	—	720	SE	5	1200		43	12	
EP	3	607	55	350290	8110090	0-5000	71	87	CG	—	1120	NE	15	2400			12	
EP	4	933	55	337490	8129060	0-5000	72	88	CG	6-4	80	SW	5	2500		59	12	
EP	9	185	55	354440	8106960	0-5000	72	88	CG	—	710	E	20	1650		63	12	
EP	18	143	55	311100	8169830	0-5000	73	87	CG	—	1100	W	5	2500			12	
EP	19	750	55	368800	7954780	0-5000	75	87	CG	—	620	SE	10	2000			12	
EP	29	650	55	345710	8059260	0-5000	75	87	AV	—	1200	SE	15	2700			12	
EP	30	144	55	293260	8199380	0-5000	76	88	CG	—	980	W	5	1500			12	
EP	31	755	55	375510	8061530	0-5000	76	88	SM	6-0	80	S	5	4000			12	
EP	32	TR 14	54	752330	8479700	0-5000	75	87	SM	2-8	450	SW	5	2000			12	
EP	33	452	55	348100	8088250	0-5000	76	88	BV	—	720	—	0	1400		52	12	
EP	34	755	55	369440	8074860	0-5000	76	88	AL	—	380	SW	5	4000			12	
EP	35	TR 55	55	322210	8190920	0-5000	77	87	SM	—	230	SE	10	2900			12	
EP	37	679	55	660835	7649387	0-5000	77	87	BV	—	920	SE	5	2400			12	
EP	38	194	55	338220	8073460	0-5000	77	87	AV	—	1000	SE	10	1800			12	
EP	40	144	55	297320	8198970	0-5000	78	88	CG	—	800	N	10	1300			12	
EP	41	NP	55	333200	8215260	0-5000	77	87	AL	—	15	SE	5	3500			12	
EP	42	CL	55	745020	8590560	0-5000	77	87	AL	—	30	SE	10	2200			12	
EP	43	194	55	333560	8085620	0-5000	78	88	AV	—	1120	S	20	2000			12	
EP	44	194	55	295120	8205880	0-5000	80	88	CG	—	880	NW	5	2500			12	